



Biosorption of Copper (II) Ions by Eclipta Alba Leaf Powder from Aqueous Solutions

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ABSTRACT

The removal of heavy metals from industrial wastewater is of great concern as heavy metals are non-biodegradable, toxic elements that cause serious health problems if disposed of in the surrounding environment. The present study, Karisalangkani (Eclipta Alba) leaves were used for the adsorption of heavy metals like copper (Cu (II)) ions. The bio sorbent was characterized using SEM and BET analysis. The bio sorption experiments are conducted through batch system. The operating parameters studied were initial metal ion concentration, adsorbent dosage, initial solution pH, contact time and effect of temperature. Adsorption equilibrium is achieved in 30 min and the adsorption kinetics of Cu (II) is found to follow a pseudo-second-order kinetic model. Equilibrium data for Cu (II) adsorption are fitted well by Langmuir isotherm model. The maximum adsorption capacity for Cu (II) ions is estimated to be 9.2 mg/g at 25 °C. The experimental result shows that the materials have good potential to remove heavy metals from effluent and good potential as an alternate low cost adsorbent. Due to their outstanding adsorption capacities, Eclipta Alba is excellent sorbents for the removal of copper (II) ions.

Keywords: *Eclipta Alba, heavy metal, adsorption, pseudo first order kinetic.*

1. INTRODUCTION

Industrial wastewater treatment for the removal of pollutants such as heavy metal ions remains a worldwide challenge. Heavy metal ions discharged with industrial effluents usually find their way into receiving water bodies such as rivers, lakes and streams increasing their pollution. Copper is one of the heavy metals most toxic to living organisms and

one of the most widespread heavy metals in the environment [1]. Although some trace metals are considered as essential plant nutrients. Most heavy metals are of considerable health and environmental concern because of their toxicity and bio accumulative behavior [2-4]. Pb^{2+} , Cu^{2+} , Fe^{3+} and Cr^{3+} are especially common metals that tend to accumulate in organisms, causing numerous diseases and disorders [5]. Copper is an essential trace element for human metabolism, however, it is toxic at concentrations above 2mg/L [6]. Ingestion of large doses irritate the mucous membranes, damage the hair and can cause necrotic changes in the liver and kidneys [7]. Copper is present in wastewaters coming from energy industries (such as PILCAM in Cameroon), the production of certain pesticides and in the manufacture of contraceptives in medicine. Removal of metal ions from wastewater in an effective manner has become an important issue. Efficient methods for the removal of metals have resulted in the development of new separation techniques. Precipitation, ion-exchange, flocculation, adsorption, electro-chemical processes, electro-dialysis, nano-filtration and reverse osmosis are commonly applied for the treatment of wastewater [8]. However, these methods are either inefficient or expensive when heavy metals exist in low concentrations [9]. Additionally, these methods may also affect the generation of secondary wastes, which are difficult to treat. Adsorption is an alternative technology in which increased amount of study has been focused because of cost effectiveness; local availability & technical feasibility for the removal of heavy metal ions from the wastewater [10]. In recent years, bio sorption has been suggested as being cheaper and more effective than chemical or physical

technologies. Low cost, high efficiency, minimization of chemical and biological sludge are the most important advantages of bio sorption technique. Moreover, bio sorbent regeneration and metal recovery is also possible. The mechanism of binding of metal ions by adsorbents may depend on the chemical nature of metal ions (species size and ionic charge), the type of biomass, environmental conditions (pH, temperature, ionic strength) and existence of competing organic or inorganic metal chelators. Natural materials that are available in large quantities or certain waste products from agricultural operation may have the potential as inexpensive sorbents.

The aim of the present study is to investigate the bio sorption of copper (II) on *Eclipta Alba* leaves powder. The effect of time, initial concentration of the metal, solution pH and adsorbent dosage on the bio sorption are also studied. The experimental data were fitted to Langmuir, Freundlich and Temkin equations, to determine which isotherm most closely correlates with experimental data. The first-order Lagergren, pseudo-second-order and second-order equations were adopted to test the experimental data.

2. MATERIALS AND METHODS

2.1. Chemicals

All the chemicals used were of analytical reagent grade. Copper nitrate trihydrate ($\text{Cu}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$) was used for preparation of stock solutions. Hydrochloric acid and Sodium hydroxide were used to adjust the solution pH. Distilled water was used throughout the experimental studies.

2.2. Biosorbent collection and preparation

The *Eclipta Alba* leaves (EA) were collected from Botanical garden present in C. P. A. College, Bodinayakanur. The EA were washed thoroughly with distilled water and boiled to remove colour and impurities. They were dried in a hot air oven at 105°C and ground to a uniform size of 0.8 mm. So prepared EA particles were used for further studies



Fig.1. Eclipta Alba leaves

2.3. Stock solution of metals

A stock solution (1000 mg/l) of Cu (II) was prepared dissolving 3.802 g of copper nitrate trihydrate in DD water. Working solution was prepared by diluting stock solution with distilled water.

2.4. Determination of Copper (II)

The copper ion concentration in the liquid phase was determined spectrophotometrically by making a complex with sodium diethyl dithio carbamate.

10 ml of the Cu (II) solution was pipetted into a beaker, and to this 5ml of 25% aqueous citric acid solution was added. The pH was adjusted to 8.5 using dilute ammonia. To this, 15 ml of 4 % EDTA was added and the solution cooled to room temperature. This was then transferred to a separating funnel, and 10 ml of 0.2% aqueous sodium diethyl dithiocarbamate solution was added and shaken for 45 sec. A yellow brown colour developed in the solution. 20 ml of Chloroform was pipetted into the funnel and shaken for 30 seconds. The organic layer acquired a yellow colour. This was shaken for 15 sec and the phases were allowed to separate. The aqueous layer was removed. To the organic layer, 20ml of 5% H_2SO_4 (V/V) was added, shaken for 15 sec, and the organic phase was separated. The absorbance was measured at 435nm against a blank.

2.5. Characterization of bio sorbents

The bio sorbent were characterized using a number of techniques including: Fourier transform infrared spectroscopy (FT-IR), scanning electron microscopy (SEM) and Brunauer-Emmet-Teller (BET) surface area analysis.

2.6. Adsorption experiment

Batch adsorption experiments of copper was carry out to determined the adsorption capacity of *Eclipta Alba* at different metal concentrations ranging from 10 to 80 mg/L and a fixed amount (100 mg/L) of *Eclipta Alba* in order to calculate the adsorption constant using different isotherms. 300 mL of different concentration of Cu (II) solutions ranging from 100–500 mg/L were used. The *Eclipta Alba* (100 mg/L) was added to flasks and agitated at 25°C and 1000 rpm for 180 min. The initial and final concentrations of the solutions were measured were determined by UV – Visible spectrophotometer at the maximum adsorption wavelength and the adsorption capacities of the adsorbent were calculated. After equilibrium was attained, the metal uptake capacity for each

sample was calculated according to a mass balance on the metal ion using equation (1):

$$q_e = \frac{(C_0 - C_e)V}{m} \quad \text{----- (1)}$$

Where m is the mass of adsorbent (g), V is the volume of the solution (L), C_0 is the initial concentration of metal (mg L^{-1}), C_e is the equilibrium metal concentration (mg L^{-1}) and q_e is the metal quantity adsorbed at equilibrium (mg/g). Experiments were carried out at different initial pH values. The initial pH of the solution was adjusted with either HCl or NaOH. The percent removal of metals from the solution was calculated by the following equation (2).

$$\% \text{removal} = \frac{C_0 - C_e}{C_0} \times 100 \quad \text{----- (2)}$$

Where C_0 (mg/L) is the initial metal ion concentration and C_e (mg/L) is the final metal ion concentration in the solution.

3. RESULTS AND DISCUSSION

3.1. SEM analysis

The surface morphology studies of bio sorbent were showed octahedral rod shaped crystals as shown in Fig. 2.

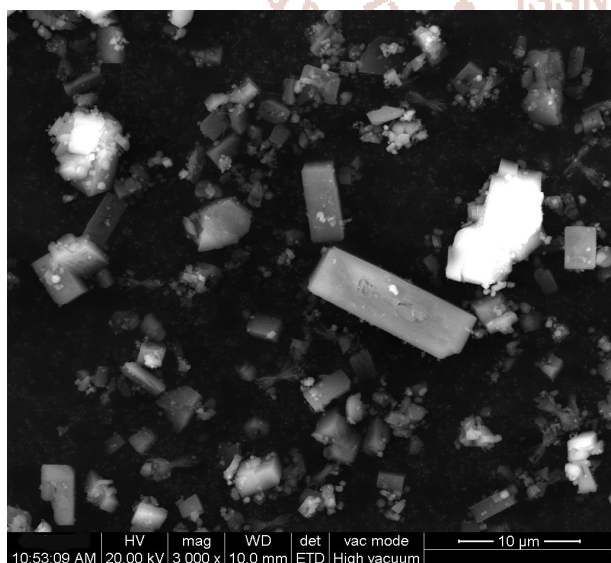


Fig.2. SEM image of EA

3.2. BET surface area analysis

The metal -bio sorption capacity, surface properties and active functional groups of any bio sorbents could be explored from surface area characterization.

Surface properties of EA were examined through BET. The surface properties of EA was measured shown in Fig. 3. These bio sorbents have noticeable micro pore and mesopore area with a magnitude of $21.17 \text{m}^2/\text{g}$. All pores of EA are mesopore (100%) and total pore volume was $0.03 \text{cm}^3/\text{g}$. In addition, the mean micro pore and mesopore sizes of the EA's powder were found to be 8.71 and 40.18 \AA , respectively, suggesting that this bio sorbent has fallen within the region of mesopore based on IUPAC-classification. The high pore area, pore volume and pore sizes indicated the numerous binding sites on the EA surfaces.

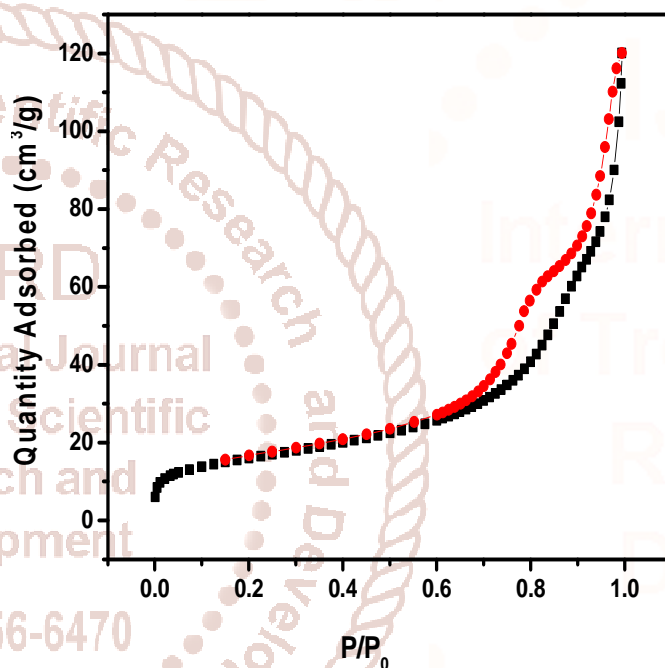


Fig.3. BET surface area of EA

3.3. Effect of initial metal ion concentration

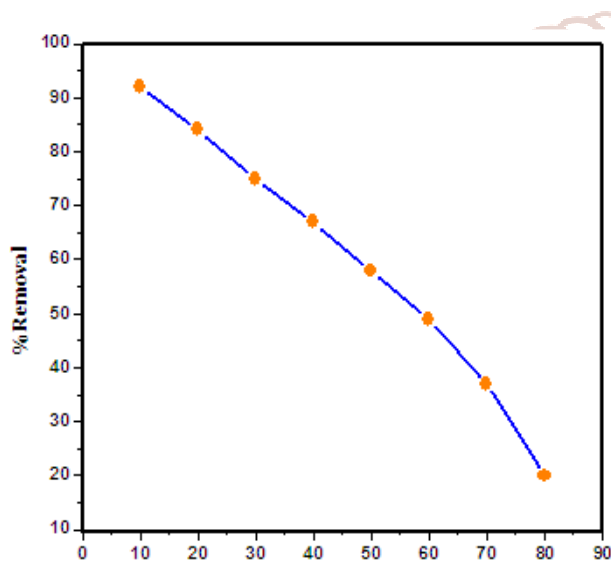
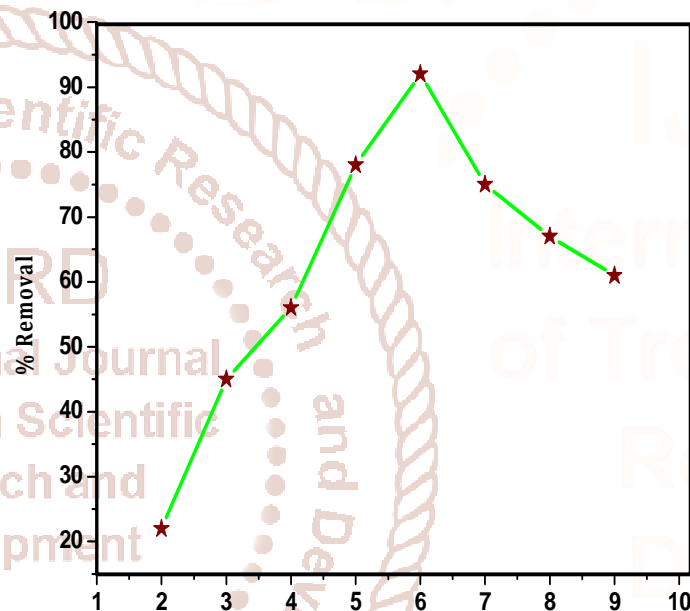
The initial concentration of Cu (II) provides an important driving force to outweigh all mass transfer resistance of metal between the aqueous and solid phases. Removal of Cu (II) for various initial concentrations (10 to 80 mg/L) by EA (100 mg) at 180 min contact time and at pH 6 has been depicted in the Fig. 4. The percentage Cu (II) adsorbed decreased 92 % to 20 % with increase in initial concentration from 10 mg/L to 80 mg/L . More adsorption capacity at higher initial metal concentration may be due the presence of more metal ion at increasing initial metal concentration. It reflects that the resistance to mass transfer reduces therefore the removal percentage decreases at higher metal concentration [11]. Based on the experimental values, 10 mg/L Cu (II) concentration is selected as the optimized value for further use.

Table-1

Initial metal ion concentration (mg/L)	% Removal
10	92
20	84
30	75
40	67
50	58
60	49
70	37
80	20

Table-2

pH	% Removal
2	22
3	45
4	56
5	78
6	92
7	75
8	67
9	61

**Fig.4. Effect of initial metal ion concentration (mg/L)****Fig.5. Effect of pH**

3.4. Effect of pH

pH influences the surface charge of the adsorbent, the degree of ionization and the species of adsorbate. So the pH of the aqueous solution is an important controlling parameter in the heavy metal adsorption process [12]. The effect of pH on the adsorption of Cu (II) by EA was shown in Fig. 5. The effect of solution pH on the adsorption of metal ions onto EA was evaluated in the pH range of 2 to 9. The percentage removal of Cu (II) increased with an increase in pH. The equilibrium adsorption capacity was observed at pH 6 for Cu (II) ions. Therefore, at low pH values (pH<2.0) EA showed very low tendency for uptake of Cu (II) ions due to protonation of its functional groups or competition of H⁺ with metal ions for binding sites. As the pH increased, there were fewer H⁺ ions present in the solution and consequently more negatively charged sites were made available and this facilitated greater metal ions uptake by electrostatic attraction.

3.5. Effect of contact time

The impact of contact time on adsorption of Cu (II) can be seen from Fig.6. It is clearly depicted in the figure that the rate of adsorption is very high initially. But at later stages, the rate of adsorption decreases. The concentration of Cu (II) ions nearly remains constant after 180 minutes. It is chiefly because of the saturation of these active sites on the EA which do not permit furthermore adsorption to occur. This can be explained by the fact that initially, the quantity of sites on the surface is very large which allows adsorption to take place very easily. But with the passage of time, the active sites get saturated thereby reducing the rate at which adsorption occurs. As a result, the extent of copper removal decreased with increase in contact time, which is dependent on the number of vacant sites on the surface of resin. Based on these results 180 min was considered as the optimum time for the rest of the experiments [13, 14].

Table-3

Contact time (min)	% Removal
0	20
30	37
60	49
90	58
120	67
150	75
180	92

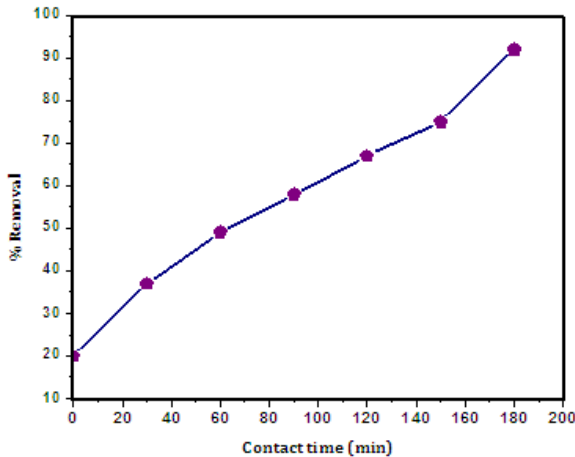


Fig.6. Effect of contact time (min)

3.6. Influence of initial adsorbent dosage

The amount of adsorbent dosage was varied in the given range 100 mg/L to 500 mg/L. It was observed from the graphs that increasing the dosage increases the % removal of Cu (II). It can be attributed to the increase in adsorbent sites for more adsorption of the Cu (II) at the fixed 10 mg/L [15].

Table-4

Asorbent dose (mg/L)	% Removal
100	22
200	44
300	58
400	75
500	92

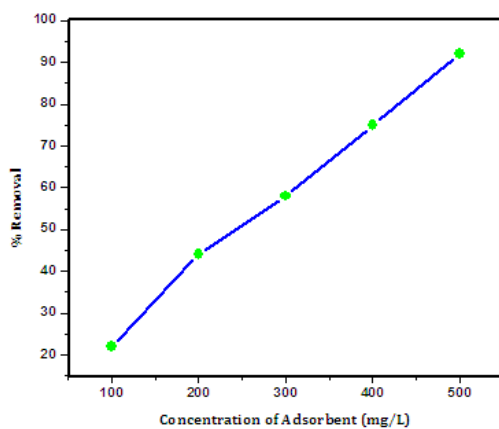


Fig.7. Effect of adsorbent concentration (mg/L)

3.7. The study of adsorption isotherm models

Among various plots employed for analyzing the nature of adsorbate-adsorbent interaction, adsorption isotherm is the most significant. The results of adsorption studies of Cu (II) at different concentrations ranging from 10 to 80 mg/L on a fixed amount of adsorbent are expressed by three of the most popular isotherm theories viz., Freundlich, Langmuir and Temkin isotherms. These isotherm equations are as follows:

Langmuir

$$\frac{C_e}{q_e} = \frac{1}{q_0 b} + \frac{C_e}{q_0}$$
 ----- (3)

Freundlich

$$\log q_e = \log K_f + \frac{1}{n} \log C_e$$
 ----- (4)

Temkin

$$q_e = RT/b \ln K_t + RT/b \ln C_e$$
 ----- (5)

Where C_e (mg L^{-1}) is the equilibrium concentration of Cu (II) in solution, q_e (mg g^{-1}) is the equilibrium adsorption capacity of EA, q_0 (mg g^{-1}) is the maximum adsorption capacity of EA for monolayer coverage, b is related to the energy of adsorption (L/mg), K_f and n are the Freundlich constants. K_f is related to the adsorption equilibrium constant, n corresponds to the number of active sites in the adsorbent required for metal ions to adsorb. Where b is the Temkin constant related to the heat of adsorption (J mol^{-1}) and K_t is the Temkin isotherm constant (L g^{-1}). The slope and intercept of linear plots of C_e/q_e versus C_e yielded values of $1/q_0$ and $1/(b \cdot q_0)$ for Equation (3). The slope and intercept of linear plots of $\ln q_e$ versus $\ln C_e$ for Equation (4) produced values of $1/n$ and $\ln K$.

In this part, the results and adsorption isotherm constants of the most common isotherms including Freundlich, Langmuir and Temkin have been illustrated in Table 5. Isotherm is assumed to be the most important parameter in designing adsorption systems and describes the relation between adsorbate concentration and adsorption capacity of an adsorbent. Regarding to Table 5 and values of linear regression coefficients. The amounts of R^2 in Langmuir isotherm for Cu (II) adsorption by EA were higher than the amount of R^2 for other isotherms. So it can be said with certainty that Cu (II) adsorption by EA follows this isotherm. Langmuir model estimated that the maximum amount of Cu (II) adsorption

capacity (q_{max}) by EA was 9.2 mg/g. K_L is a constant that increases when the size of adsorbent molecules increased [17]. If dimensionless parameter of $R_L=0$, the adsorption is irreversible and if $0 < R_L < 1$, the adsorption is Favorable and if $R_L=1$, the adsorption is linear and if $R_L > 1$, then the adsorption is unfavorable. According to the results of the Langmuir isotherm for adsorption of Cu (II), the amounts R_L for EA was between 0 and 1, so Cu (II) adsorption by EA is Favorable [18].

Table-5 the Isotherm constants for Cu (II) adsorption onto EA

Isotherm	Parameters	Cu (II)
Langmuir	R^2	0.9799
	q_e	9.012
	R_L	0.2784
	b	0.31079
Freundlich	R^2	0.9045
	n	2.3191
	K_f	1.965
Tempkin	R^2	0.8565
	b (kJ/mol)	5.248
	K_t (L/mg)	17.89

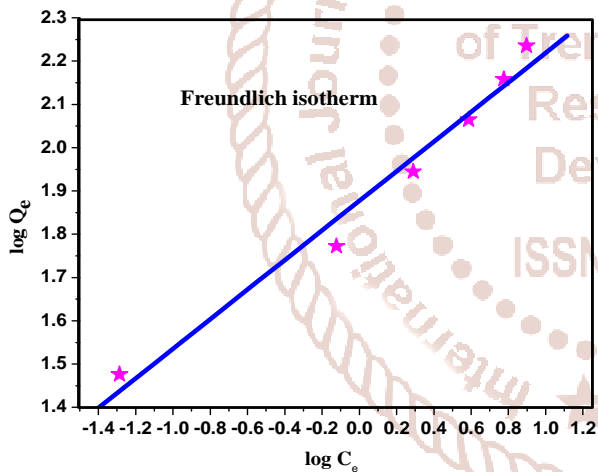


Fig.8. Freundlich isotherm

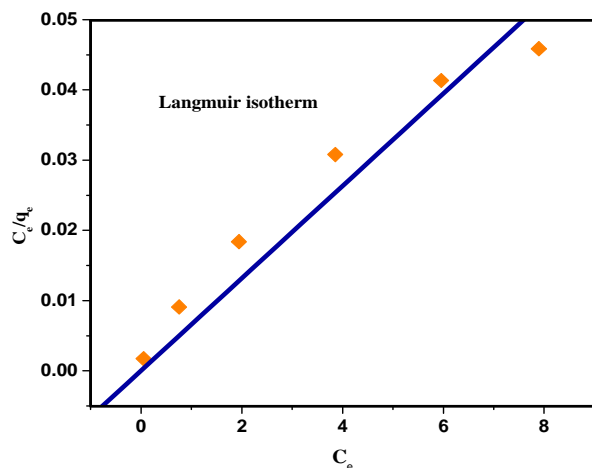


Fig.9. Langmuir isotherm

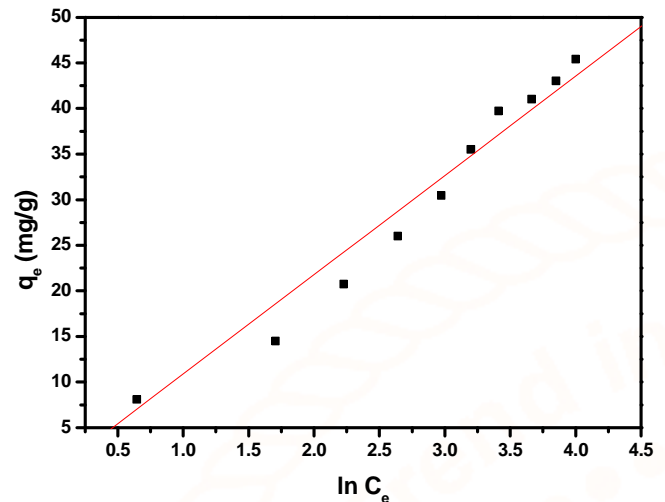


Fig.10. Temkin isotherm

3.8. Adsorption kinetics

Several models can be used to express the mechanism of Cu^{2+} adsorption onto an EA bio sorbent. The kinetic models of adsorption were determined with two most commonly used and known rate equations: pseudo first order (the most reliable kinetics equation suitable only for the rapid initial phase Equation 6) and pseudo second order (for predicting the kinetics behavior of chemical sorption as a rate controlling step Equation 7) models are expressed as

$$\log (q_e - q_t) = \log q_e - \frac{K_1}{2.303} \cdot t \quad \text{----- (6)}$$

$$\frac{t}{q_t} = \frac{1}{K_2 q_e^2} + \frac{1}{q_t} \cdot t \quad \text{----- (7)}$$

Where q_e and q_t are the amounts of metals adsorbed ($mg\ g^{-1}$) at equilibrium and at time t (min) respectively. K_1 (min^{-1}) is the rate constant of pseudo-first order. The linear relation among the value of $\log (q_e - q_t)$ versus t (Fig. 11) reveals the applicability of this model. K_2 and the equilibrium adsorption amount (q_e) were measured from the slope and intercept of the plot of q_t versus t (Fig. 12). Where k_2 ($g/(mg\ min)$) is the rate constant of pseudo-second order kinetic model. h is known as initial sorption constant (Table - 6). The high values of R^2 were recorded for the Pseudo second order compared to pseudo-first order model. This indicates that the adsorption kinetic was better represented by the pseudo-second order model for all metal ions. Therefore, the overall rate of the metal ions adsorption process appears to be controlled by the chemical process via ion exchange and/or complex process [19-21].

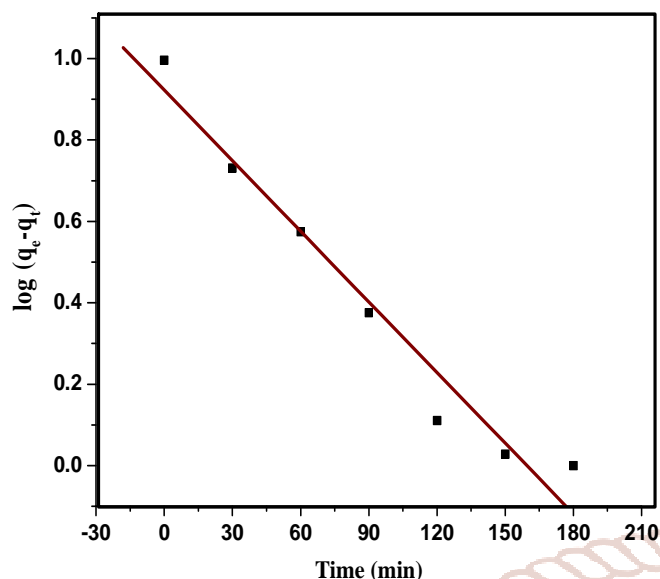


Fig.11. Pseudo first order kinetics

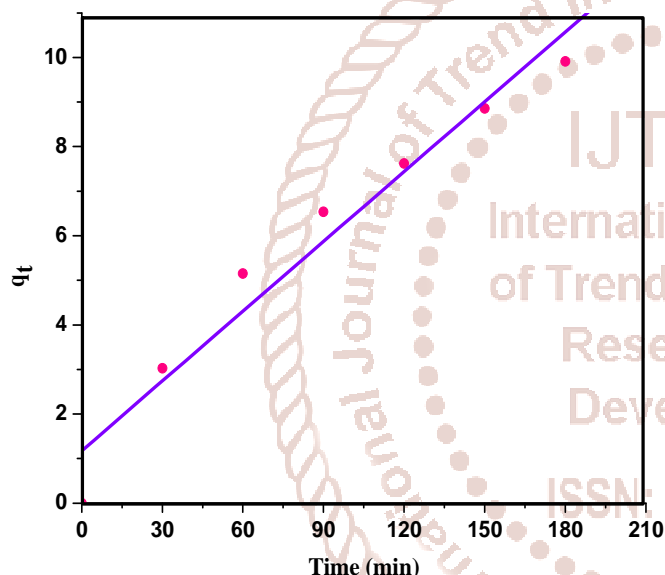


Fig.12. Pseudo second order kinetics

Table-6: Different kinetic model parameters

Models	Parameters	Cu (II)
Pseudo-first order model	$q_{e, exp}$ (mg/g)	9.2
	R^2	0.85
	k_1 (min^{-1})	0.0207
	$q_{e, cal}$ (mg/g)	0.657
Pseudo-second order model	R^2	0.984
	h (mg/g min)	0.1034
	k_2 (g/mg min)	0.0834
	$q_{e, cal}$ (mg/g)	9.35

4. CONCLUSION

The potential of natural Eclipta Alba as a low cost material for the removal of copper from synthetic metal solutions was studied and the bio sorbent were successfully characterized using SEM and BET

technique. A number of experiments were performed in order to determine this potential; equilibrium adsorption capacity, % removal of metal ions. The maximum adsorption capacities observed for 10 mg/L of Cu (II). A maximum of 92 % metal ion removal was observed for copper. Various parameters like pH, initial ion concentration, adsorbent dosage, and contact time were studied successfully. The maximum removal of metal ions at 10 mg/L of Cu(II) with 100 mg/L of adsorbent dosage. The study indicates that the Eclipta Alba is a good candidate as a low cost adsorbent to be used for the removal of heavy metals from wastewater.

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